

Extinction potential from invasive alien species

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33 **Abstract**

34

35 Biological invasions pose significant threats to biodiversity, while impacting ecosystem
36 services, human health, and cultural heritage. Despite these far-ranging effects, their
37 impacts are generally underappreciated by both the public and policymakers, resulting in
38 insufficient management and inadequate conservation outcomes. Recognizing the gap in
39 effective quantitative measurement tools, we introduce the Extinction Potential Metric (EPM)
40 and its derivative, EPM for Unique species (EPM-U; adjusted for phylogenetic uniqueness)
41 to quantify the ecological damage caused by invasive alien species (IAS). These metrics
42 estimate the number of current and projected extinct species within a 50-year horizon under
43 a business-as-usual scenario due to specific IAS.

44

45 We applied EPM and EPM-U to assess threats to native terrestrial vertebrates from IAS,
46 examining impacts on 2178 amphibians, 920 birds, 865 reptiles, and 473 mammals. The
47 analysis identified that damage mostly stems from a limited number of IAS, notably two
48 pathogenic fungi affecting amphibians (up to 380 equivalent extinct species) and primarily
49 cats (139 equivalent extinct species) and rats (50 equivalent extinct species) impacting other
50 groups, through mechanisms such as predation, disease, and reduced reproductive success
51 in birds.

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53 The proposed metrics not only provide a standardised measure of ecological impacts but are
54 sufficiently versatile to be tailored for specific spatial and temporal scales or taxonomic
55 groups. Furthermore, EPM could serve as a model for developing unified indicators to
56 monitor global biodiversity targets, such as those defined in the Kunming-Montreal Global
57 Biodiversity Framework (GBF), by assessing the ecological effects of various individual and
58 combined anthropogenic stresses. Also, EPM and EPM-U could support the enforcement of
59 Target 6 of the GBF, by establishing lists of IAS requiring urgent prevention and control.

60 Thus, EPM and EPM-U offer critical tools for improving the management of biological
61 invasions and enhancing global conservation strategies.

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64 **Keywords:** Conservation, Extinction, Biological invasion, Ecological impact, Metric,
65 Phylogeny, Red List

66

67 **Introduction**

68

69 Biological invasions are a major threat to biodiversity and cause impacts across all aspects
70 of the natural and human world, including reduced ecosystem services, increased human
71 health hazards, and loss of indigenous and cultural practices (Roy et al., 2024). The number
72 of newly introduced alien species has been steadily increasing since the 1800's (Seebens et
73 al., 2017), and the number of established alien species is expected to increase by 36%
74 between 2005 and 2050 under a business-as-usual scenario (Seebens et al., 2021). The
75 subset of these species that have negative impacts, so-called invasive alien species (IAS),
76 have contributed to 60% of known species extinctions, cost hundreds of billions of \$US each
77 year (increasing four-fold every decade and reaching \$423 billion in 2019), and adversely
78 impact quality of life in 85% of cases (Roy et al., 2024). Despite these figures, the impacts of
79 biological invasions remain underestimated and poorly understood by both the public and
80 policy-makers, leading to hindered public support and participation, inadequate management
81 actions, and hampered conservation outcomes (Courchamp et al., 2017).

82

83 The recent *InvaCost* database, which compiles the economic costs of IAS (including
84 management- and damage-related costs), provides a quantitative metric documenting the
85 economic impacts of IAS (i.e. monetary costs in a standardised currency and year; Diagne et
86 al., 2020, 2021). As such, it succeeded in facilitating public and policy awareness (Ahmed et
87 al., 2023). For ecological impacts, multiple metrics have also been proposed, primarily based
88 on impact categories. For example, the Environmental Impact Classification for Alien Taxa
89 scheme (EICAT; Blackburn et al., 2014; Hawkins et al., 2015) categorises the maximum
90 impact an IAS has had globally into four categories above minimal (undetected) impact, and
91 has been adopted by the IUCN (IUCN, 2020b, 2020a). These EICAT scores range from
92 impacts on individual fitness to the local extinction of species and irreversible changes in

93 community composition (see Bernardo-Madrid et al., 2022 for an assessment of the
94 consistency of seven impact classification schemes).

95

96 However, to our knowledge, there is no fully quantitative, continuous impact metric that
97 would allow for comparison of the ecological impacts of IAS across taxa and scales. Such a
98 metric would allow for more granular impact comparisons among contexts and taxa with
99 more detailed analyses and predictions, such as spatial and temporal trends of derived
100 indicators, quantitative links with invasion drivers and alien species traits, or disentangling
101 the demographic and per-unit or per-capita components of impact, therefore allowing for
102 more targeted management actions (Latombe et al., 2022). Since managing biological
103 invasions is often the most effective conservation action (Langhammer et al., 2024), robust
104 and comparable quantitative metrics would permit prioritisation of action as requested from
105 Target 6 of the Kunming-Montreal Global Biodiversity Framework (GBF), which aims to
106 “eliminate, minimise, reduce and or mitigate the impacts of invasive alien species on
107 biodiversity and ecosystem services” (CBD, 2022). It would also underpin indicators
108 appropriate for risk identification and prioritisation efforts to meet global biodiversity targets
109 (Butchart et al., 2005; Vicente et al., 2022). The lack of such standardised metrics capable of
110 capturing the severity of IAS impacts to ecosystems and their services has been a major
111 impediment, not only to action, but also to effective communication between scientists,
112 stakeholders and policy-makers.

113

114 Here, we propose the Extinction Potential Metric (EPM) as a measure of impact of individual
115 IAS. The EPM of an IAS can be conceptualised as the number of species that have already
116 been led to extinction plus those that are expected to go extinct within a 50-year time frame
117 under a business-as-usual scenario because of this IAS. It can also be seen as the
118 equivalent number of extinct species, allowing for the comparison of IAS with different
119 current levels of impacts on multiple native species, from complete extinction to population
120 decrease or range contraction. The EPM uses the existing IUCN Red List of Threatened

121 Species framework (IUCN, 2024), quantifying the threats posed by IAS to native fauna to
122 compute a risk of extinction within 50 years for each native species threatened by IAS.
123 Moreover, since native species extinctions are unequal in terms of evolutionary
124 distinctiveness among taxa, we introduce an additional metric, namely EPM-U, representing
125 the number of 'unique' native species that have and would go extinct within a 50-year time
126 frame. Species uniqueness accounts for the evolutionary history of the impacted native
127 species. EPM-U therefore compares IAS with impacts on phylogenetically related vs
128 unrelated species. Below, we present the EPM and EPM-U metrics and their computations
129 in detail. We showcase both metrics with a proof of concept on native terrestrial vertebrate
130 species threatened by IAS (amphibians, reptiles, birds and mammals). We also show how
131 EPM and EPM-U can be used to identify which impact mechanisms are most detrimental to
132 native species. EPM and EPM-U complement other methodologies to measure the impact of
133 IAS, and in particular EICAT (Blackburn et al., 2014; Hawkins et al., 2015).

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135

136 **Method**

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139 **Data: Native species impacted by identified invasive alien species**

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141 We utilised the IUCN Red List data (IUCN, 2024) to catalogue terrestrial vertebrate species
142 impacted by IAS, i.e. those associated with Threat 8.1 (Invasive non-native/alien
143 species/diseases). We identified 2178 amphibians, 920 birds, 865 reptiles, and 473
144 mammals in categories Least Concern (LC), Near-Threatened (NT), Vulnerable (VU),
145 Critically Endangered (CR), Extinct in the Wild (EW), Extinct (EX), and Data Deficient (DD)
146 for which at least one IAS was listed as a threat. DD species were discarded from the
147 analyses. Invasive organisms threatening these native species were identified at the species

148 level when possible (73%), but some were only identified at the genus (7%), family (1%), or
149 order and class (0.6%) level. For 19% of the threats, the invasive organism was not
150 identified and was recorded as “Unspecified species”. We refer to entries in the database as
151 IAS hereafter for simplicity, and to genus, family, or order when specifically addressing these
152 categories. For each of the native species previously identified, we also distinguished
153 between species endemic to islands and other species, using data from Marino et al. (2022).

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155 We also extracted information on the scope (proportion of population affected) and severity
156 (speed of population decline) of the IAS threat to these native species from the database.

157 Those with over 50% of their population witnessing a very rapid (>30% over 10 years or
158 three generations; whichever is the longer), rapid (20–30% over 10 years or three
159 generations; whichever is the longer), or slow and significant (<20% over 10 years or three
160 generations; whichever is the longer) decline due to IAS were deemed to experience a high
161 impact, while the other species were considered to experience a low impact (definitions from
162 IUCN & CMP, 2012a). When neither the extent nor the severity was available, impact level
163 was considered as ‘not available’ (NA). In addition, we extracted the mechanism by which
164 each native species was affected by IAS, using the IUCN Red List stress classification
165 scheme (IUCN & CMP, 2012b).

166

167 In addition to tabular information summarising the threat details on native species described
168 above, the IUCN Red List website provides more detailed textual descriptions of the threats
169 affecting native species, including from IAS, interlinking data with the Global Invasive
170 Species Database implemented by the IUCN Invasive Species Specialist Group.

171 Examination of these textual descriptions revealed that species included in the description
172 are often missing from the tabular information. To assess the effects of this difference of
173 information content, we manually examined the description for the 473 mammals in the LC,
174 NT, VU, CR, EW, and EX categories, and generated an extended database including all
175 species mentioned both in the text and in the table. Comparison between the extended

176 dataset and the tabular dataset for mammals showed that the tabular information could miss
177 up to 10 native species impacted by a given IAS with a mean value of 0.89 ± 1.12 (Supp. Fig.
178 1).

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180 **Computing the probability of extinction of each native species**

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182 To convert IUCN Red List categories into probabilities of extinction, we followed the EDGE2
183 and FUSE approaches (Gumbs et al., 2023; Pimiento et al., 2023) (Fig. 1). These
184 approaches first assign the following species extinction probabilities within 50 years based
185 on their IUCN Red List category: 0.06 for LC, 0.12 for NT, 0.24 for VU, 0.49 for EN, and 0.97
186 for Critically Endangered CR. EX and EW were attributed a probability of 1. To mitigate
187 potential biases resulting from the discrete categorisation of species extinction risk, we then
188 applied probabilistic draws for VU to CR species. These five IUCN Red List categories were
189 first transformed into ordinal values from 1 to 5, and a quartic distribution was then applied to
190 link these ordinal values and the extinction probabilities using the `polyfit` function from the
191 `pracma` R package V 2.4.4 (Borchers, 2023) (Figure 1). For each species, a value was
192 randomly chosen from a uniform distribution spanning the range of its ordinal category value
193 ± 0.5 , using the `runif` function in R. We then computed its probability of extinction based
194 on the previously established quartic distribution (Figure 1). This process was replicated 10
195 times for each native species, generating a set of 10 extinction probabilities per species
196 according to their IUCN Red List category, enabling us to generate confidence intervals and
197 account for uncertainty.

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199 **Translating the extinction risks of native species into future number of species** 200 **extinct due to IAS**

201

202 We calculated the extinction potential of each IAS by summing the probabilities of extinction
203 (including EX and EW species) for all native species impacted by this specific IAS. However,
204 the contribution of an IAS to the extinction risk or to the past extinction of a native species
205 can vary. According to the IUCN Red List, the impact of IAS on a native species can be
206 classified as high or low (Fig. 1), but these data can also be missing (hereafter NA). To
207 account for these different levels of impact of IAS on the decline of native species population
208 globally (low, high, or NA—as described above), we divided the probability of extinction of
209 native species by a factor 2 for low impacts. For IAS whose impacts were classified as NA
210 on some native species, we performed two sets of analyses to account for different
211 uncertainty scenarios. First, assuming that NAs potentially reflect very small impacts, we
212 excluded the corresponding native species in the computation of EPM values, providing a
213 conservative value. As a result, 10 EPM values were computed for each IAS, based on the
214 10 extinction probabilities generated for each native species, as described above, adjusted
215 for the level of impact of IAS.

216

217 Second, assuming NAs may simply represent a gap in how impacts were reported rather
218 than a level of impact, we randomly selected between low and high impact categories, and
219 repeated this process 100 times to assess the uncertainty linked to data deficiency. As a
220 result, for each original set of 10 probabilities of extinction generated for each native species,
221 we calculated 100 scores when the impact level of an IAS was NA, corresponding to their
222 probability of extinction adjusted for the impact level. Consequently, to each of these IAS
223 corresponded 1000 sets of extinction probabilities for the native species they impacted (10
224 extinction probabilities drawn from the quartic function \times 100 potential impact levels). For
225 each IAS, we therefore obtained a distribution of the number of native extinct species. We
226 then computed the EPM for all native species together, but also for each native taxonomic
227 group separately (i.e. each IAS had an impact distribution for amphibians, birds, mammals,
228 and reptiles).

229

230 **Considering unspecified IAS**

231 As some IAS were only identified at the genus and family levels in the IUCN Red List
232 database, we also used information on unspecified species to compute the maximum
233 possible EPM value for an IAS based on the available data, by assuming unspecified
234 species included all species in the reported genus or family. In other words, we considered a
235 given IAS to be responsible for all the threats for which it is identified, but also for all the
236 threats imposed by unspecified IAS of the same genus or family. For example, the maximum
237 EPM of *Rattus rattus* was computed by aggregating the EPM of *Rattus rattus*, the EPM of
238 Unspecified Rodentia and the EPM of Unspecified *Rattus*. The maximum EPM of the other
239 *Rattus* species, i.e. *Rattus norvegicus*, and *Rattus exulans*, was computed in the same way.
240 Conversely, doing so implies that native species impacted by unspecified *Rattus* species
241 were considered as affected by all *Rattus* species (i.e. *Rattus rattus*, *Rattus norvegicus*, and
242 *Rattus exulans*).

243 **Accounting for evolutionary distinctiveness**

244

245 Evolutionary distinctiveness is an important component of biodiversity that can be
246 incorporated into our approach by following an approach similar to that used to compute the
247 EDGE2 metric (Gumbs et al., 2023). The EDGE2 score of a species represents the number
248 of millions of years of evolutionary history lost if a species goes extinct. It is the product of
249 the heightened evolutionary distinctiveness (ED2) of a species and of its risk of extinction
250 (GE2, for global endangerment). ED2 accounts for the phylogenetic distance of a species
251 from other species (its “raw” evolutionary distinctiveness ED), but also the extinction risk of
252 these other species. A species with an intermediate ED whose phylogenetically related
253 species are endangered will therefore have a higher ED2, as the risk of losing the whole
254 clade is then higher.

255

256 To account for evolutionary history in measuring the impact of IAS, we propose an additional
257 metric, the Extinction Potential Metric for Unique species due to an IAS (EPM-U), which
258 incorporates ED2 scores into EPM. As ED2 values are only available for mammals (Gumbs
259 et al. 2023), we only computed EPM-U of IAS on native mammals. To do so, we propose to
260 use the maximum ED2 score of all mammals to determine the most unique species in the
261 database, with a value of 1. We then rescaled all ED2 values by dividing them by this
262 maximum value, to generate a “uniqueness score” for each impacted native species in [0,1].
263 In addition, Gumbs et al. (2023) account for uncertainty in the calculations of ED2 values
264 and provide a median value, with the corresponding lower and upper interquartile ranges for
265 each species. We sampled 10 ED2 values per native species from a normal distribution
266 where the mean is the median ED2 value and the standard deviation is the lower
267 interquartile interval divided by 1.35. These 10 ED2 values for each native species were
268 rescaled between 0 and 1 after dividing them by the maximum ED2 value. We then
269 computed the number of unique extinct species due to an IAS as the sum of the probabilities
270 of extinction for each native species impacted by the IAS multiplied by their uniqueness
271 score. Following the same methods as for the EPM, we generated 100 scores per IAS.

272

273 **Linking impact strength to impact mechanisms**

274

275 We assessed the importance of the 11 impact mechanisms listed in the IUCN Red List
276 through which IAS affect native species, using EPM values. For each of these 11
277 mechanisms—competition, ecosystem conversion, ecosystem degradation, hybridisation,
278 inbreeding, indirect ecosystem effects, reduced reproductive success, skewed sex ratios,
279 species disturbance, species mortality, and species stresses—we extracted the native
280 species impacted by any IAS (identified or not). We then computed the corresponding EPM
281 score based on the IUCN Red List category of the native species, the severity, and the
282 scope of the threats, generating an EPM score for each impact mechanism following the

283 same framework as detailed above. This was done for all native species and for each
284 taxonomic group separately.

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287 **Results**

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289 **Global impacts of IAS on native taxa**

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291 Considering all native species for which the impact level of IAS was known, we observed
292 from the 196 IAS that median EPM values (all EPM values hereafter correspond to median
293 values) ranged from 0.02 (*Pavo cristatus*) to 89.8 (*Batrachochytrium dendrobatidis*) extinct
294 species, with a mean value of 2.4 ± 9.64 (hereafter corresponding to the standard deviation).
295 EPM values followed a log-normal distribution, with a few species having high values and
296 most having low values and 47 IAS with a value above 1 extinct species (Fig. 2.A, Supp.
297 Table 1). When including native species with an impact level defined as 'NA' (i.e. randomly
298 assigned to low or high 100 times), 357 IAS were included, and mean EPM values increased
299 by 38% with 80 IAS above 1 extinct species. The EPM values of IAS with the greatest
300 impact were particularly affected, with *B. dendrobatidis* rising from 89.8 to 380.7, *Felis catus*
301 from 87.2 to 138.8, and *Batrachochytrium salamandrivorans* from 15.1 to 88.5 (Supp. Fig.
302 2.A).

303

304 When computing EPM values of IAS for the different native taxonomic groups separately
305 (i.e. the equivalent number of extinct native amphibians, birds, mammals or reptiles) we
306 observed different rankings across taxa (Fig. 2.B, Supp. Table 2). Five IAS had an EPM
307 higher than 1 extinct amphibian species. Two pathogenic fungi were particularly impactful, *B.*
308 *dendrobatidis* (the pathogen causing chytridiomycosis in frogs; EPM = 89.8.0), followed by
309 *B. salamandrivorans* (EPM = 15.1). Twenty-six IAS had an EPM above 1 extinct bird

310 species, and three IAS above 10: *Felis catus* with (EPM = 40.1), *Rattus rattus* (EPM = 27.2),
311 and *Canis familiaris* (EPM = 10.7) (Fig. 2.B). Five IAS had an EPM score above one extinct
312 mammal species, but one was particularly impactful: *Felis catus* (EPM = 22.6) (Fig. 2.B).
313 Finally, 15 IAS had an EPM score above 1 extinct reptile species, and 3 had a score above
314 10: *F. catus* (EPM = 25.0), *Wasmannia auropunctatus* (EPM = 12.2), and *Herpestes*
315 *auropunctatus* (EPM = 11.3) (Fig. 2.B). When we included native species with an impact
316 defined as 'NA', EPM values increased particularly for amphibians, for which the number of
317 IAS with a score above 1 rose from 5 to 19 (a 280% increase). The number of IAS above 1
318 rose by 30% for birds (34 vs 26), 100% for mammals (10 vs 5) and 60% for reptiles (25 vs
319 15) (Supp. Fig. 2.B, Supp. Table 3). Similarly, we observed substantial increases in EPM
320 values for the most impacting IAS for amphibians (*B. dendrobatidis* EPM = 380.5 and *B.*
321 *salamandrivorans* = 87.0) but also for reptiles (*F. catus* EPM = 51.8, *R. rattus* EPM = 21.3,
322 and *H. auropunctatus* EPM = 19.8) and to a lesser extent for birds (*F. catus* with EPM =
323 54.1, *R. rattus* EPM = 32.3, and *C. familiaris* EPM = 14.4) and mammals (*F. catus* EPM =
324 31.7) (Supp. Fig. 2.B, Supp. Table 4).

325

326 Results were similar for maximum EPM scores considering unspecified genera or families,
327 with *Rattus* species seeing the main increase, due to many impacts being reported for
328 unspecified *Rattus* and rodents in the IUCN Red List database (Supp. Fig. 3). *Mus musculus*
329 and *H. auropunctatus* had a substantive increase in their score when considering
330 unspecified rodents and unspecified *Herpestes*.

331

332 Insular endemic species accounted for $82.0 \pm 25.7\%$ of the previously computed EPM scores
333 across all native species. Specifically, the impact of IAS was highest for island endemic
334 amphibians and reptiles, with mean changes of $-92.4 \pm 27.2\%$ and $-88.7 \pm 16.6\%$, respectively.
335 The effect was slightly less pronounced in birds, at $-80.6 \pm 27.8\%$, and was the weakest in
336 mammals, at $-62.0 \pm 37.0\%$ (Fig. 3). Adding native species with 'NA' impact did not broadly
337 affect these results (Supp. Fig. 4).

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Impacts of IAS on mammal species: tabular vs. textual information and number of unique extinct species

When considering the extended database comprising the tabular (high, low, and 'NA' threats) and textual information (only 'NA' threats), 29 IAS had an EPM score of more than 1 extinct mammal species (Fig. 4). *F. catus* had the largest impacts (EPM = 64.2), followed by *C. familiaris* (EPM = 30.6), *R. rattus* (EPM = 20.0), and *Vulpes vulpes* (EPM = 19.3). When considering the evolutionary distinctiveness of impacted mammals, the same 4 species had the highest EPM values (*F. catus* = 3.9, *C. familiaris* = 2.2, *V. vulpes* = 1.9, and *R. rattus* = 0.98) (Supp. Fig. 6.B). The evolutionary distinctiveness of impacted mammals had little effect on the ranking of the four most impactful IAS, but had important effects on rankings for other IAS (Fig. 5).

Impacts caused through different mechanisms

When considering all taxa, 'Species mortality' was associated with the largest EPM value of 396.9, more than three times the second and third highest EPM value from 'reduced reproductive success' (EPM = 122.5) and 'ecosystem degradation' (EPM = 109.6) (Fig. 6.A). Furthermore, 'Species mortality' was the most consistent mechanism, showing a significant association with higher EPM scores for amphibians, birds, mammals, and reptiles (Fig. 6.B). For native birds, it was closely followed by 'reduced reproductive success,' and for native birds and reptiles, 'ecosystem degradation' also had a substantial effect (Fig. 6.B). Adding native species with 'NA' impact did not affect 'species mortality' which stayed largely the most impactful mechanism with EPM = 1018.7, but 'ecosystem degradation' became the second most important mechanisms with EPM = 235.3 above 'reduced reproductive success' EPM = 158.4 when considering all taxa (Supp. Fig. 6).

366

367 **Discussion**

368

369 We introduced and applied a new metric, the Extinction Potential Metric, which computes the
370 ecological impacts of IAS as the equivalent number of native species that IAS have
371 contributed to driving to extinction. The EPM uses the IUCN Red List assessments to
372 compute extinction probabilities of native species impacted by IAS within 50 years, in
373 addition to those species already extinct. A quantitative EPM value is then attributed to each
374 IAS based on the number of species it is expected to drive to extinction, allowing for detailed
375 comparison of the impacts of different IAS. We also introduced an alternative metric, EPM-U,
376 the Extinction Potential Metric for Unique species, to account for differences in evolutionary
377 history of impacted native species, and complement the information provided by EPM.

378

379 **Most impacts on native tetrapods result from a few IAS**

380

381 Birds, mammals, and reptiles were all mostly impacted by cats (*F. catus*), followed by dogs
382 (*C. familiaris*) and rats (especially the black rat, *R. rattus*, but also other rat species, often
383 unspecified or including *Rattus norvegicus* and *Rattus exulans* for birds), foxes (*V. vulpes*),
384 the small Indian mongoose (*H. auro punctatus*), and domestic goats (*Capra hircus*). All these
385 species belong to the IUCN list of 100 of the worst invasive species (Lowe et al., 2000;
386 Luque et al., 2014). Results are nonetheless different for amphibians, which are mostly
387 affected by two invasive pathogenic fungi, *B. dendrobatidis* and *B. salamandrivorans* (Fig.
388 2), with the former also appearing in the 100 of the worst list. The number of amphibian
389 species affected by these two species is higher than the number of all tetrapod species
390 affected by any other IAS (see also Bellard, Genovesi, et al., 2016). An important part of
391 these impacts is recorded for species that are endemic to islands (Fig. 3), indicating the
392 disproportionate impacts IAS have on species that are potentially evolutionarily distinct

393 within insular habitats (Millien, 2006). In addition, for most IAS, the numbers of impacted
394 native species in each Red List category are of the same order of magnitude (Fig. 7, Supp.
395 Fig. 7). This means that EPM scores are mostly driven by past extinctions (EX and EW
396 categories) or by decreases in native species populations and ranges that have driven them
397 to the brink of extinction (CR category), rather than by expected future extinctions (LC to VU
398 categories).

399

400 The evidence presented above confirms that controlling a limited number of IAS could
401 effectively prevent and mitigate the impacts of biological invasions on biodiversity. This is
402 consistent with the conclusions of the IPBES assessment that highlighted that, with sufficient
403 resources, political will and long-term commitment, preventing and controlling IAS are
404 attainable goals that will yield significant long-term benefits for nature (Roy et al., 2024).
405 Also, the results of our analysis support the need to identify priority IAS for prevention and
406 control, as requested from Target 6 of the GBF.

407

408 Considering the evolutionary distinctiveness of impacted native mammals, did not
409 fundamentally change the ranking of the most impactful IAS (Fig. 5), but changed the score
410 distribution, with a more gradual increase in EPM-U compared to EPM values (Supp. Fig. 5).
411 Evolutionary distinctiveness, and more generally phylogenetic diversity, is a core component
412 of conservation as it captures multiple concepts (Gumbs et al., 2023; Lean & Maclaurin,
413 2016; Winter et al., 2013). Evolutionarily distinct species may be considered to have a high
414 intrinsic value on the basis of their uniqueness. Phylogenetic diversity (maximised by
415 conserving the most unique species) has also been advanced as a proxy for functional
416 diversity and evolutionary potential, whose maximisation is important for ecosystem services
417 and for maintaining the benefits humans draw from nature, although these claims have been
418 questioned (Mazel et al., 2018; Winter et al., 2013). The difference in EPM and EPM-U score
419 distribution has implications for the design of management actions and IAS prioritisation.
420 Although assessing the risk caused by IAS is insufficient to prioritise which IAS to manage,

421 which should also account for management feasibility and costs (Robertson et al., 2021), it is
422 an essential component of this process. Here, using EPM scores alone would give a greater
423 weight to the few species with a disproportionately high score, whereas EPM-U scores
424 would suggest considering a wider range of species.

425

426 “Species mortality” is the main mechanism through which IAS impact native species (Fig. 6).
427 This occurs through predation for mammals, birds, and reptiles, and through two diseases
428 for amphibians. Ranking of the other mechanisms varied substantially across the four
429 taxonomic groups. Interestingly, birds were the only group for which reduced reproductive
430 success was an important mechanism, ranking second. However, this is likely due to the fact
431 that IAS consume native bird eggs before hatching, which is therefore similar to predation in
432 the sense that this is linked to direct resource consumption. Aggregating the EPM scores of
433 these two impact mechanisms for native birds would generate a profile similar to that of the
434 other taxonomic groups. Other important but secondary, non-lethal impact mechanisms
435 include ecosystem degradation, species disturbance, and competition. The predominance of
436 direct predation over other impact mechanisms suggests that appropriate IAS management
437 requires the removal of interactions between IAS predators and their prey. It may also
438 suggest that increased attention is needed towards more subtle threat mechanisms which
439 are less conspicuous than those driving direct mortality. While lethal control of predators
440 may seem to be the most obvious approach, it may not be effective or even achievable,
441 especially on the mainland (García-Díaz et al., 2021). IAS populations can recover quickly,
442 preventing eradication success, and lethal control may need to be performed at levels that
443 are not attainable with current techniques, or are financially unsustainable to see positive
444 effects. Efficacy is also context-dependent, with studies on various native taxa showing
445 effects of IAS control ranging from positive to negligible and even negative effects (García-
446 Díaz et al., 2021). Context-specific alternative approaches to lethal control have been
447 proposed, including IAS exclusion, for example using fences (D. Smith et al., 2020), nest
448 cages to prevent egg consumption (Buzuleciu et al., 2015; R. K. Smith et al., 2011),

449 translocation of native species (Miskelly & Powlesland, 2013), or prey training and
450 harnessing natural selection to combat prey naivete (Moseby et al., 2016), although their
451 efficacy has been questioned. Nevertheless, IAS eradications have had high success rates
452 on islands (Spatz et al., 2022), which contributed the majority of extinction risk in those
453 habitats.

454

455 The results presented here should be taken with some level of caution. First, the IUCN Red
456 List, despite being a major conservation resource, is not fully comprehensive, and not all
457 native nor invasive species have been assessed or are up-to-date (Cazalis et al., 2022,
458 2023). This is exemplified by the large shares of IAS with impacts classified as 'NA' towards
459 particular native species in the current study. The resource is also more likely to report
460 impacts from IAS when these have been long standing, and often uses observed and
461 inferred impact through expert opinion rather than comprehensive, published evidence (e.g.
462 Gula et al., 2023; Van der Colff et al., 2021). The figures we present are therefore
463 conservative. The IUCN Red List is also of limited use in identifying future problematic IAS,
464 given time lags to invasion and impact (Essl et al., 2011). Finally, and importantly, the IUCN
465 Red List data is open access and can easily be extracted from the IUCN Red List website as
466 text files containing data in a tabular format. However, examination of more detailed textual
467 information provided on the webpage of each assessed native mammal revealed that this
468 tabular information was incomplete, with up to 10 IAS missing in the list of threats (Supp.
469 Fig. 1). Computing EPM scores for the extended database comprising the textual information
470 revealed that impacts were largely under-estimated, even for the most impactful species,
471 and that ranking differed for many species. We therefore advocate for an in-depth curation of
472 the IUCN Red List database, to harmonise the information content on IAS threats on native
473 species between tabular and textual information, to ensure consistency across taxa.
474 Fostering knowledge on IAS impacts on native species is also paramount to improve these
475 figures and develop an even more accurate application of EPM.

476

477 **EPM as a measure of ecological impacts**

478

479 The new metric we propose provides a bespoke quantitative measure of ecological impact of
480 IAS and has many advantages compared to existing, mostly qualitative metrics. Using
481 probabilities, the EPM links impacts at the population and species level. It uses species as a
482 unit, which is the fundamental unit of conservation, as exemplified by the IUCN Red List of
483 Threatened Species—arguably the most influential global conservation tool to date. As the
484 IUCN Red List is open access, EPM assessments are transparent, readily applicable and
485 reproducible. The EPM score of an IAS can be interpreted as the number of species that it
486 will have driven to extinction in 50 years under a business-as-usual scenario, including past
487 and future extinctions. Thus, EPM is easily conceptualised by policy-makers and the civil
488 society, and will therefore foster communication about the impacts of biological invasions
489 and the design of management actions, such as native conservation prioritisation.

490

491 In addition, EPM offers the possibility to be temporally and spatially downscaled. Red List
492 statuses are ideally updated every few years for each species, which will enable us to
493 generate estimates and track temporal changes in the impacts of IAS. EPM, as presented
494 here, assesses ecological impacts of IAS at the global scale, because it is the scale for
495 which the IUCN most commonly provides Red List categories and with the most complete
496 level of details. However, Red List categories can also be assessed at national and regional
497 levels (IUCN, 2012). EPM could therefore also be spatially explicit if the same level of
498 information on the status and threats on native species is assessed at multiple spatial
499 scales. This downscaling is particularly pertinent given that the impacts of IAS can differ
500 substantially among populations occurring in different regions over time, undermining the
501 efficacy of global species ‘watch lists’.

502

503 As demonstrated in our analyses, EPMs can be computed for different groups of impacted
504 native species, allowing for the capture of multiple dimensions of impacts caused by IAS,
505 rather than generating a single value or category for an IAS. Here, we showed that IAS
506 impacts varied across four taxonomic groups of native tetrapods. In addition, although we
507 applied the EPM at the IAS level, it would be straightforward to upscale the metric to the
508 genus level or beyond, or for any group of IAS of interest. For example, here we also
509 quantified the importance of each mechanism through which IAS impact native species, by
510 computing EPM scores after grouping all IAS impacting native species through these threat
511 mechanisms. Although the most important impacts occur through direct mortality, it also
512 revealed that other, non-lethal impacts (e.g. competition or species disturbance, such as the
513 disruption of critical lifecycle stages) or changes in the environment (e.g. ecosystem
514 degradation) can also have important impacts. Doing so therefore enables us to capture the
515 broader ecological impacts of biological invasions for different components of ecosystems,
516 and how they cascade to *in fine* affect native species survival.

517

518 Note, however, that native species impacted by biological invasions are often threatened by
519 multiple IAS simultaneously, as well as other anthropogenic pressures, and their effects can
520 combine non-additively. Disentangling the relative impacts of different IAS on a single native
521 species is not possible with the current granularity of IUCN data. It is consequently not
522 possible to assess if an IAS could lead a native species to extinction in the absence of
523 another IAS (e.g. cats are often associated with red foxes in Australia, but have led many
524 native species to extinction by themselves in other regions). In the absence of a clearly
525 identifiable rule to determine the contribution of each IAS, here we took a conservative,
526 objective and reproducible approach, attributing the same probability to cause extinction to
527 different IAS affecting a native species with the same level of impact. This follows
528 precautionary principles, an approach already used by other impact classification schemes,
529 including EICAT (Kumschick et al., 2024), to avoid irreversible impacts or high costs to the
530 environment and society that would result from inaction.

531

532 The versatility of EPM offers avenues to derive indicators assessing the ecological impacts
533 of biological invasions to inform international policy initiatives, in particular Target 6 of the
534 GBF. This framework is lacking appropriate indicators that would allow tracking of policy
535 performance regarding this target (Henriksen et al., 2024). Such indicators should capture
536 spatial and temporal changes, be species-specific and allow for identification of the
537 mechanisms through which IAS impact native species (Henriksen et al., 2024), which are all
538 readily included in EPM. Different indicators have been developed, but are either qualitative
539 (e.g. Genovesi et al., 2012), categorical (e.g. Kumschick & Nentwig, 2010), do not provide
540 information on the IAS and their impact mechanisms (e.g. Butchart, 2008; Butchart et al.,
541 2005), or focus on species extinction (i.e., extreme 'end-points') rather than combining
542 multiple levels of earlier-stage impacts on different native species (e.g. Bellard, Cassey, et
543 al., 2016). Indicators derived from EPM could be combined with other indicators of impacts
544 (Van der Colff et al., 2021) but also with trends or projections in IAS numbers and ranges,
545 trends in introduction and spread mechanisms, and trends in policy responses (McGeoch et
546 al., 2006, 2010; Rabitsch et al., 2016), to provide a comprehensive description of the issue
547 and analysis of its drivers.

548

549 Finally, the EPM could be used to assess the ecological impacts of other anthropogenic
550 pressures listed as threats under the IUCN Red List classification scheme (e.g. residential &
551 commercial development, agriculture & aquaculture, pollution, climate change & severe
552 weather, etc., as per the IUCN Red List Threat Classification scheme V3.3; IUCN & CMP,
553 2012a). Doing so could lead to a set of harmonised indicators to assess progress towards
554 different targets of the GBF (e.g. Target 7: Reduce Pollution to Levels That Are Not Harmful
555 to Biodiversity; CBD, 2022). It could also permit analyses of synergies among global change
556 drivers, by identifying which processes combine to create the highest threats, compared to
557 where they occur individually. Indeed, specific native species are often impacted by multiple
558 anthropogenic pressures simultaneously, which can reinforce each other.

559

560 **Conclusion**

561

562 We introduced the Extinction Potential Metric, a new quantitative metric to assess the
563 ecological impacts of invasive alien species. The EPM computes the equivalent number of
564 native species driven to extinction by an IAS using the IUCN Red List categories to assign
565 probabilities of extinction within 50 years. Doing so enables us to incorporate IAS impacts at
566 the population and species levels into a single metric, and also to account for evolutionary
567 distinctiveness between impacted native species. Applying this metric to native terrestrial
568 vertebrates shows that a few IAS disproportionately affect native species populations
569 through direct mortality, suggesting that controlling a limited number of IAS could effectively
570 prevent and mitigate the impacts of biological invasions on biodiversity. The metric is
571 versatile and transparent, and can be applied to different groups of native species and IAS,
572 to permit prioritisation of conservation action and to provide a detailed and more
573 comprehensive assessment of IAS ecological impacts than existing approaches. We argue it
574 can be the basis to derive spatio-temporal indicators that will improve our capacity to assess
575 the efficacy of global and regional conservation policies.

576

577

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586

587

588 **Data availability**

589

590 The data on which this paper is based are freely available on the IUCN Red List website
591 (www.redlist.org). A list of the native and invasive alien species used in the analyses is given
592 in the electronic supplementary material.

593

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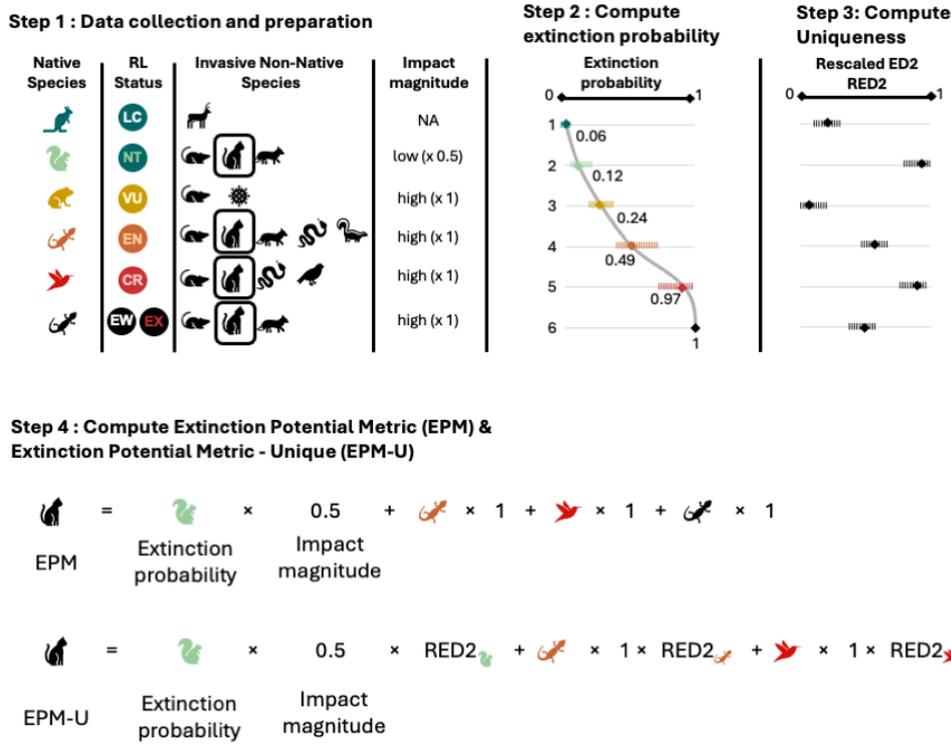
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790 **FIGURES**

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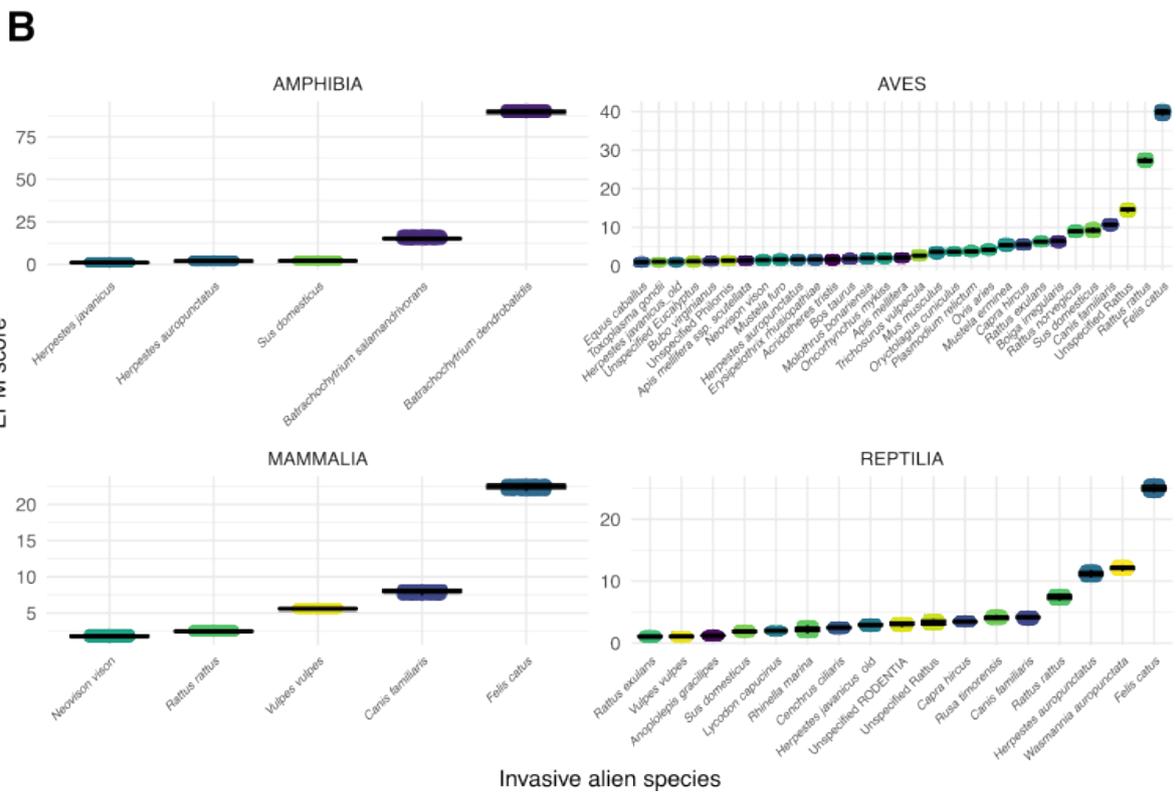
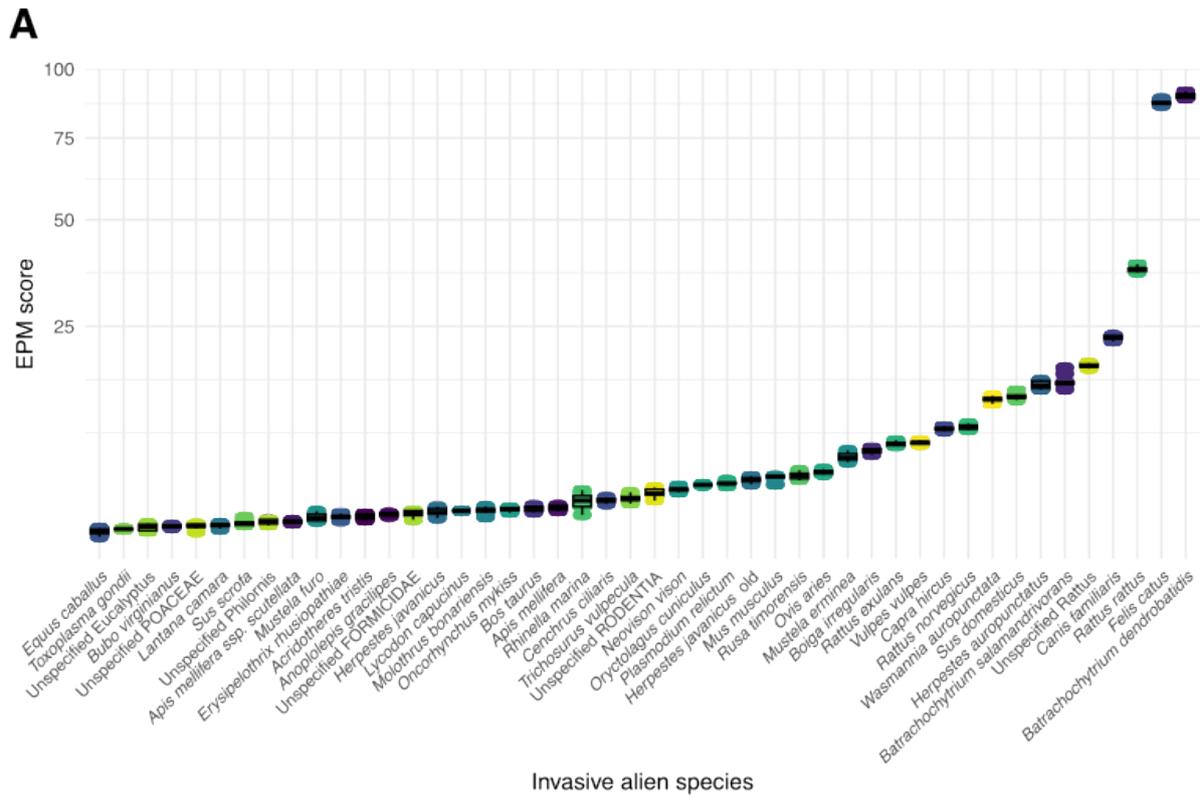
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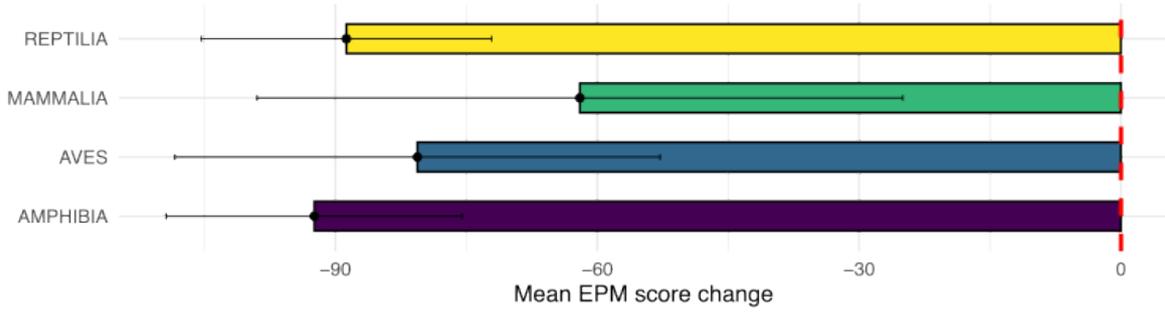
Figure 1. Workflow used to compute the Extinction Potential Metric. Species uniqueness was computed by rescaling ED2 scores from Gumbs et al. (2023); although they are indicated for all species here to show the general principle, ED2 scores were only available for native mammals. For impact magnitude, NAs were either considered as no impact (x 0) or randomly drawn as low or high impacts. Rug plots depict random draws along the quartic function for extinction probabilities and between lower and upper interquartiles from Gumbs et al. (2023) for ED2 scores.



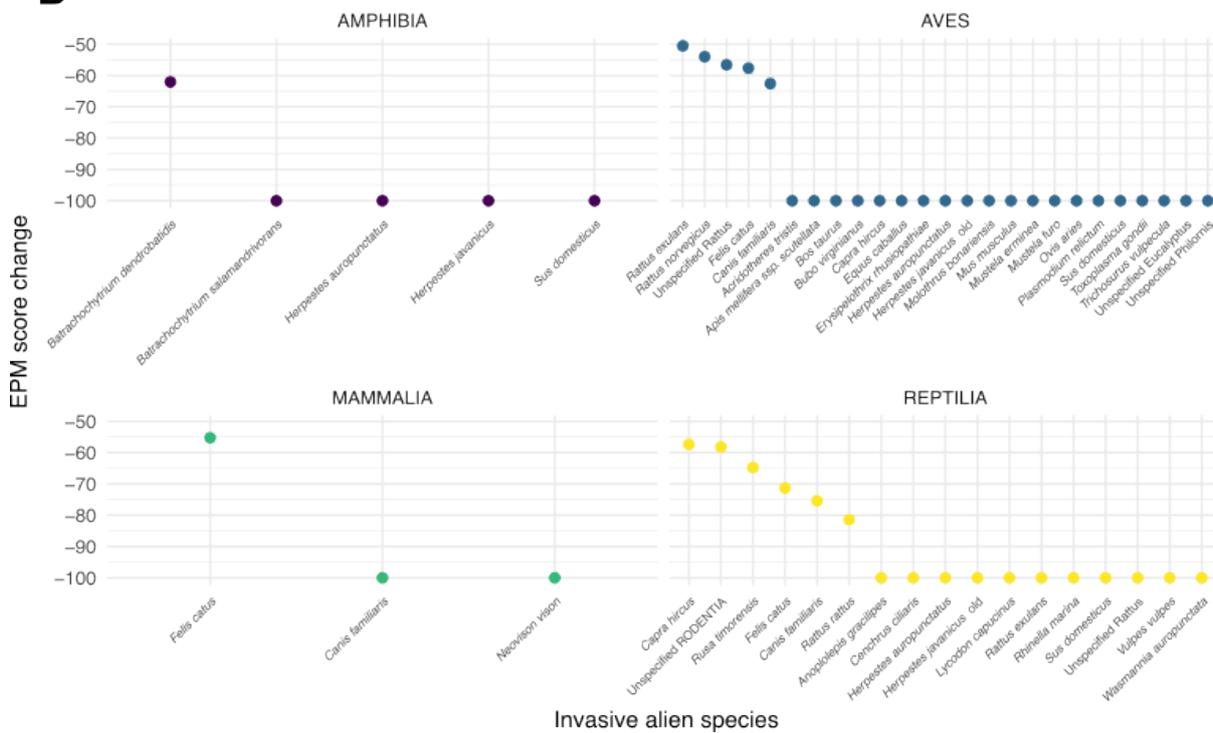
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Figure 2. Extinction Potential Metric (EPM) scores of invasive alien species (IAS) impacting native tetrapods globally. Only IAS with a score above 1 are displayed. Panel A shows EPM scores based on the impacts of IAS on all native tetrapods combined. Panel B shows EPM scores based on the impacts of IAS on each taxonomic group (amphibians, birds, mammals, and reptiles) separately.

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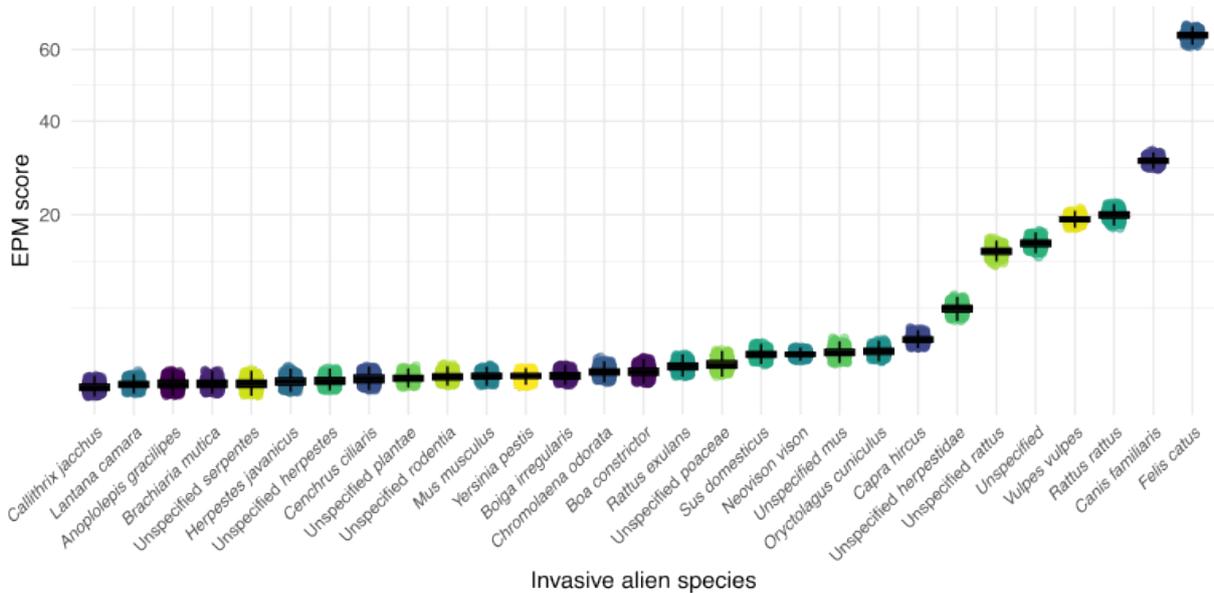
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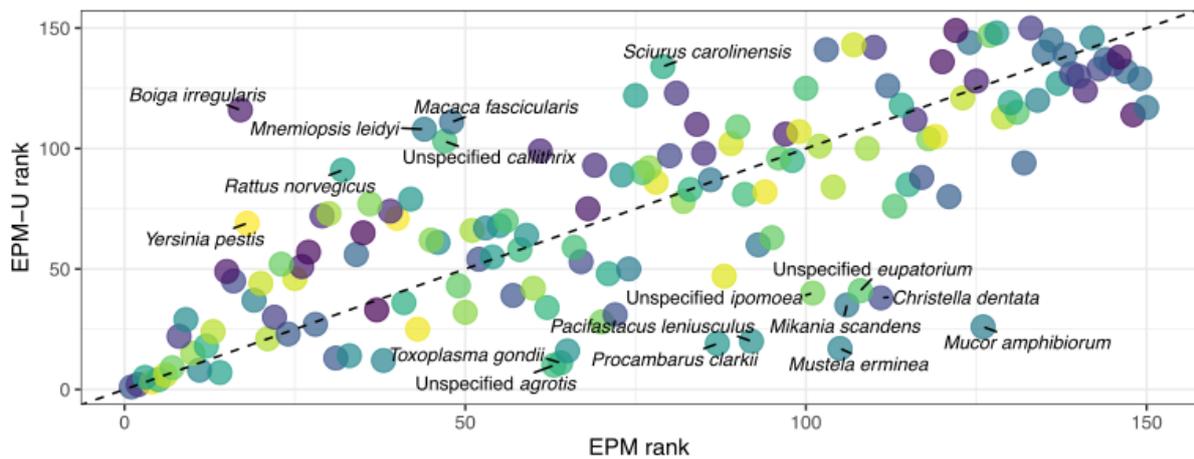
Figure 3. Contribution of insular endemic species to the EPM scores. EPM scores reflect the impact of IAS on native tetrapods globally, considering insular endemic species. Panel A shows the change in EPM scores (in %) when excluding insular endemic species for each of the four native taxonomic groups. Panel B highlights IAS with a difference in score of more than 50% when excluding insular endemic species

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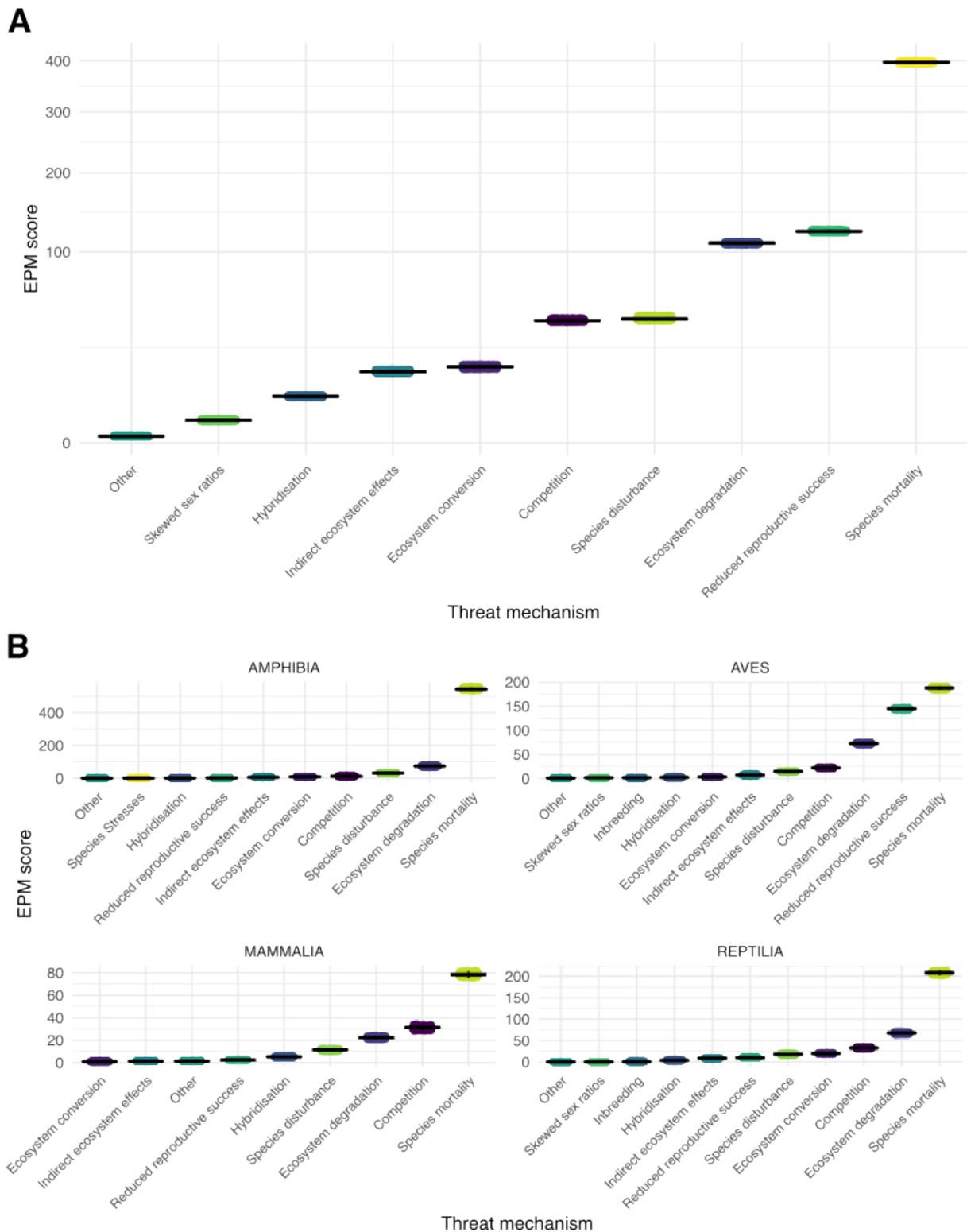
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Figure 4. EPM scores of IAS impacting native mammals globally using textual and tabular data from the IUCN Red List. EPM scores were computed using both textual and tabular data from the IUCN Red List, including all impacts on native species, even when the severity or scope of the impact was not assessed.

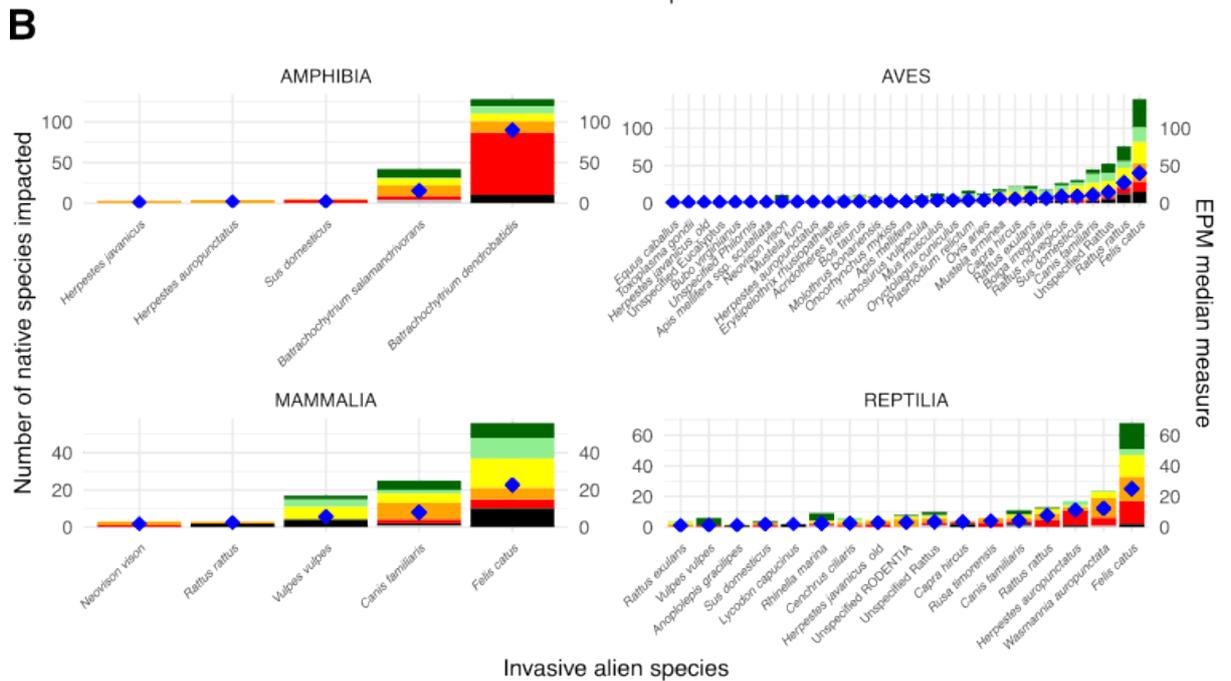
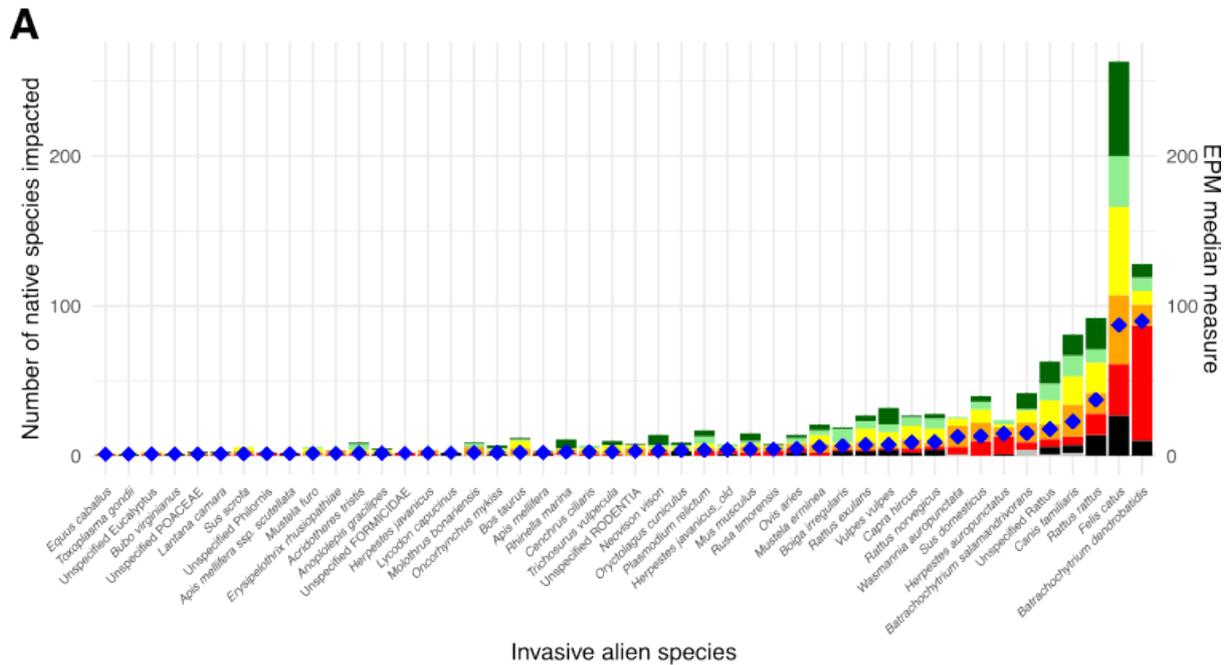


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Figure 5. Comparison of EPM-U and EPM scores of IAS according to rank. The lower an IAS's rank, the higher its score, and vice versa. Species close to the diagonal have similar ranks for both approaches. IAS whose absolute difference between the two metrics is greater than 50 are displayed with their scientific name.



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 841 **Figure 6. EPM scores of threat mechanisms impacting native tetrapods globally.** Only
 842 IAS with a score above 1 are displayed. Panel A shows EPM scores based on the impacts of
 843 mechanisms on all native tetrapods combined. Panel B shows EPM scores based on the
 844 impacts of mechanisms on each taxonomic group (amphibians, birds, mammals, and
 845 reptiles) separately.



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Figure 7. IUCN categories of native tetrapods impacted by IAS globally and their associated EPM scores. Only IAS with an EPM score above 1 are displayed. Panel A shows EPM median measure (blue diamonds) and the number of native species per IUCN category, based on the impacts of IAS on all native tetrapods combined. Panel B shows EPM median measure (blue diamonds) and the number of native species per IUCN category, based on the impacts of IAS on each taxonomic group (amphibians, birds, mammals, and reptiles) separately.